



VIA Electronic Mail (bay-delta@waterboards.ca.gov)

February 7, 2011

State Water Resources Control Board
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P.O. Box 2000
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Re: Nov 2010 SJR flow and S. Delta salinity Response

Dear Board Members:

The San Joaquin River Group Authority offers the following comments on the salinity portion of the staff presentation made by State Board staff on January 6, 2011.

Slides # 57 and 58. Staff presented a regression analysis for salinity found at Old River at Tracy in relation to salinity found at Vernalis. From information presented by the panel, the concept of using a regression analysis for this site needs to be reconsidered. Information presented in the workshop by DWR and others also show that little Vernalis flow reaches this site and that there is no relationship between the two. A simple regression does not show cause and effect.

Slides #59, 60 and 61. These slides conclude that salt loading from NPDES dischargers is minor when compared to the salt load passing Vernalis. This may be true based on the volume of flow and loads passing Vernalis compared to the volume of discharge and loads from these NPDES Permit holders however these NPDES discharge flows remain in the vicinity of Old River at Tracy and are not assimilated into the flow that moves down Old River from the main stream San Joaquin River. DWR modeling shows that little of the Vernalis flow that finds its way into Old River moves past this monitoring point near Tracy and most of the flow moves down the Grant Line Canal. The result is that the salt load from these discharges stays in the vicinity of the monitoring point (Old River near Tracy) and continues to move back and forth by tidal action. The role of NPDES dischargers and their influence on salinity measured at the Old River near Tracy monitoring point needs to be reevaluated.

Slides #64 and 65. During the SWRCB staff presentation of the Hoffman Report, it was noted that there was a potential for boron toxicity (Slide #64) and that further evaluation needs to be done (Slide #65). It should be clarified that this is a soil problem on the entire west side of the San Joaquin Valley, Delta and the Sacramento Valley. It is

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unrelated to water quality at Vernalis. At no time have the boron levels in water at Vernalis exceeded levels that would permit full use by irrigated agriculture. The entire west side of the San Joaquin River from the Delta to Bakersfield has soils that were derived from marine sediments and thus are boron enriched. This soil boron enrichment even continues into the Westside of the Sacramento Valley as shown in Yolo, Colusa, Glenn Counties' ground water wells. While soil boron in the western portion of the Delta may be limiting production in some cases, it is not related to water quality or the quality of water diverted from the San Joaquin River or Delta.

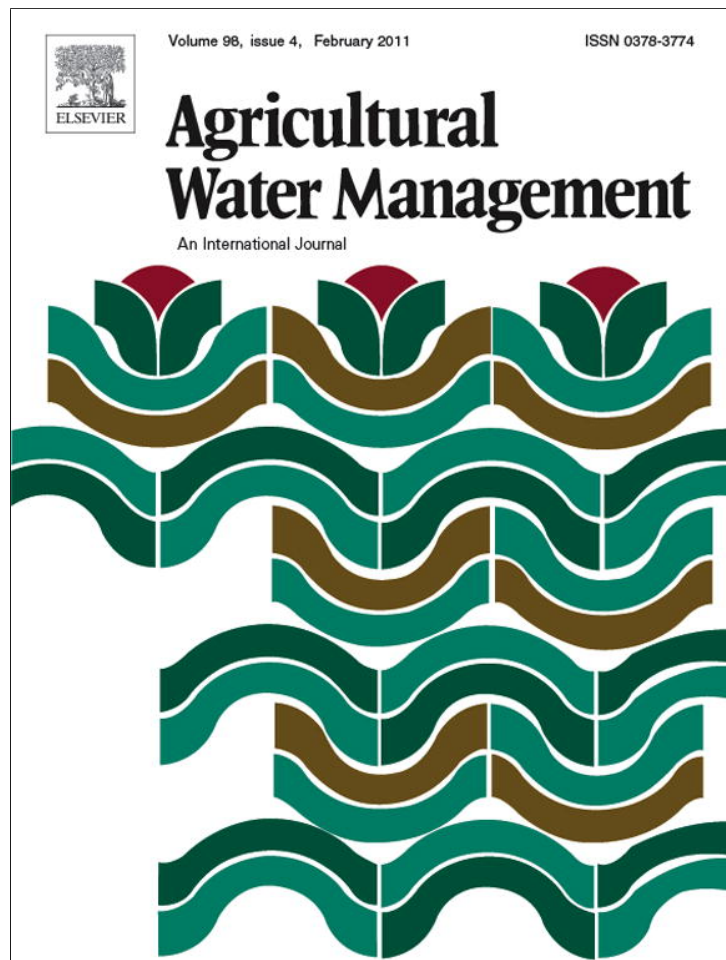
Slide # 66. The Hoffman Report recommends the use of a steady-state modeling approach for estimating salinity impacts with the present cropping pattern in the South Delta. There has been discussion that the model may be too conservative and over-estimate salinity impacts and other discussion that the use of steady-state models does not protect Delta agriculture. However, this over-estimation of impacts is often considered a margin of safety that other models do not provide. Steady-state models nonetheless remain valid, as confirmed in a recent review of the use of steady-state models in the peer-reviewed journal publication *Agricultural Water Management*. The journal article is entitled "Evaluation of soil salinity leaching requirement guidelines". The exact citation is *Agricultural Water Management* 98 (2011) 502-506 (ISSN 0378-3774). A copy of the article is attached.

Very truly yours,
O'LAUGHLIN & PARIS LLP

By:


KENNETH PETRUZZELLI

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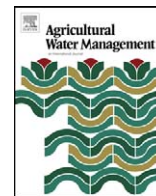
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Contents lists available at ScienceDirect

Agricultural Water Management

journal homepage: www.elsevier.com/locate/agwat

Review

Evaluation of soil salinity leaching requirement guidelines

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ARTICLE INFO

Article history:

Received 26 April 2010

Received in revised form 5 August 2010

Accepted 12 August 2010

Available online 29 October 2010

Keywords:

Irrigation

Salt leaching

Steady-state analyses

Transient-state models

ABSTRACT

Water for irrigation is a major limitation to agricultural production in many parts of the world. Use of waters with elevated levels of salinity is one likely option to meet the supply of increased demands. The sources of these waters include drainage water generated by irrigated agriculture, municipal wastewater, and poor quality groundwater. Soil salinity leaching requirements that were established several decades ago were based on steady-state conditions. Recently transient-state models have been developed that potentially can more correctly predict the dynamics of the chemical–physical–biological interactions in an agricultural system. The University of California Center for Water Resources appointed a workgroup to review the development of steady-state analyses and transient-state models, and to determine whether the current recommended guidelines for leaching requirement based on steady-state analyses need to be revised. The workgroup concludes that the present guidelines overestimate the leaching requirement and the negative consequences of irrigating with saline waters. This error is particularly large at low leaching fractions. This is a fortuitous finding because irrigating to achieve low leaching fractions provides a more efficient use of limited water supplies.

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1. Introduction

Water for irrigation is a major limitation to agricultural production in many parts of the world. Irrigated lands produce more than 40% of the world's food and account for almost 90% of the global developed water consumption (Döll and Siebert, 2002). Moreover, the increase in human population has caused an increase in water demand for urban and agricultural purposes. Use of

waters with elevated levels of salinity is one likely option to meet the supply of increased demands. The sources of these waters include drainage water generated by irrigated agriculture, municipal wastewater, and poor quality groundwater. Irrigation with saline waters requires application of extra water for leaching of salts from the root zone to prevent excessive accumulation of salts that would limit the yield potential of crops. Thus a trade-off exists between applying more water for soil salinity control and applying less water to protect groundwater quality from the leaching of pollutants such as pesticides and nitrate, and making more water available for other purposes. Advancements in irrigation technology such as micro-irrigation and sprinkler systems provide the opportunity to irrigate with very low leaching fraction (LF) values.

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However, irrigating with low leaching fractions may lead to soil salination, thereby causing reduction in crop yield.

Irrigation guidelines that were established several decades ago were based on steady-state conditions, and are still widely used to assess the suitability of water for irrigation. These guidelines were established to develop simple relationships to estimate crop yield potential from irrigation water salinity and LF. Recently though, high-speed computers have facilitated the development of transient-state models that potentially can more correctly predict the dynamics of chemical–physical–biological interactions in natural agricultural systems, thereby more accurately assess spatio-temporal changes in soil salinity during and between growing seasons. Based on computer simulation results to date, there is a growing concern that the steady-state approach may overestimate the negative effects of saline irrigation water on crop production for a range of irrigation water management conditions, as will be elaborated on in this communication.

The University of California Center for Water Resources appointed a workgroup with the charges: (1) to review the development of steady-state analyses and transient-state models, (2) to answer the question, “Do the current recommended guidelines on leaching requirements (based on steady-state analyses) need to be revised?”, and (3) if yes, draw up a new set of guidelines. This information is not only important to farmers, but also for regulatory agencies that apply or establish salinity standards for water bodies designed to protect agricultural production. This article reports the findings of the workgroup on the first two assignments.

Many articles have been published on managing saline irrigation waters, and each could be considered as providing guidelines. We chose to restrict our analyses to those presented in the Food and Agricultural Organization of the United Nations report by Ayers and Westcot (1985) that have been widely used internationally; and those of Hanson et al. (2006) published by the University of California, Division of Agriculture and Natural Resources as those promulgated by the University of California.

2. Background information

Salinity and its effects on a society have been traced back to ancient Mesopotamia (Jacobsen and Adams, 1958). Ancient history has recorded several major occurrences of salinity, but the earliest and one of the most serious occurred during 2400–1700 B.C. in what is now southern Iraq. Water was diverted from the Tigris and Euphrates rivers to irrigate fertile lands. Copious quantities of water were applied leading to water table rise and ultimate salination of the soil by capillary rise. Declining yields, abandoning salinized fields and shifts to cultivation of more salt-tolerant crops paralleled increasing soil salinity. The southern part of the alluvial plain never recovered from the decline resulting from salination. The land was converted from fertile soils to the present desert conditions.

These and other historical experiences have led to inclusion of subsurface drainage systems in irrigation projects to maintain water tables below the root zone, thereby allowing leaching of excess salts from the root zone. Since all irrigation waters contain salts and crops take up nearly pure water for transpiration (nutrients and some specific salts are taken up by the roots), salts will remain behind and concentrate in the crop root zone, thus requiring periodic leaching to move excessive salts downwards to avoid reduced crop yields. The fraction of the water amount that drains beyond the root zone relative to the amount of applied irrigation water is defined as the leaching fraction (LF).

The term “leaching requirement” (LR) has been defined as the minimum LF that is required over a growing season for a particular quality of water to achieve maximum yield of a given crop and has

a specific quantitative value. It is not simply an abbreviated way of stating that leaching is required.

A large amount of research has been conducted in the past to assess crop salt tolerance. Much of that work was summarized for more than 60 agricultural crops by Maas and Hoffman (1977). They reported salt tolerance information using two parameters: the salt tolerance threshold value and the percent yield decline per unit increase in salinity beyond the threshold value. The Maas and Hoffman (M–H) coefficients continue to provide the scientific basis for irrigation management guidelines world-wide. Coefficients were related to the average root zone electrical conductivity of the saturated soil extract (EC_e). Plants are expected to respond to the salinity of the water surrounding the root (EC_s). Since soils are typically at field capacity or at lower water content during the growing season, it has been commonly assumed that EC_s is approximately equal to $2EC_e$. Maximum yield is expected if the average root zone EC_e is equal to or less than the M–H threshold value (EC_e^*). As reported hereafter, there is a direct connection between crop salt tolerance, irrigation water salinity, and leaching requirement.

3. The meaning of steady-state and transient-state

Mathematically a steady-state flow analysis does not include a time variable; whereas, a transient-flow analysis does. Considering a flow analysis of water and solute, the water content and solute concentration at a given point remains constant with time in a steady-state system and can vary in a transient-state system. However, many reports on managing saline irrigation waters do not adhere to this strict mathematical definition of steady-state.

In fact, “true” steady-state conditions never exist in the field. Steady-state specifies that applied irrigation water is continuously flowing downwards at a constant rate, irrespective of irrigation frequency. In addition, steady-state specifies that evapotranspiration is constant over the growing season. Consequently, steady-state solutions assume that the salt concentration of the soil solution at any point in the soil profile is constant at all times. None of these is real. However, some aspects of steady-state are observed in the field. If a given irrigation regime is followed for an extended period of time, the salt concentration below the root zone does become constant with time. Thus, a steady-state condition exists below the root zone even though transient conditions exist within the root zone. The law of conservation of matter combined with the assumption that there is no chemical precipitation or dissolution can be used to calculate the “steady-state” salt concentration below the root zone. The relationship is

$$\frac{C_{dw}}{C_{iw}} = \frac{D_{iw}}{D_{dw}} \quad (1)$$

where C_{iw} and C_{dw} are the salt concentrations of the irrigation water and drainage water at the bottom of the root zone, respectively; and D_{iw} and D_{dw} are the depths of applied water and drainage water, respectively. Typically, the electrical conductivities of the irrigation and drainage waters are used for salt concentrations.

Furthermore, although the salt concentration distribution with depth in the root zone does vary with time, the pattern consistently repeats itself with time for a given irrigation regime that is followed for an extended period of time. This condition has sometimes been referred to as a steady-state condition. However, this is not a “true” steady-state condition and cannot be calculated from steady-state analysis. This condition can, however, be computed from a transient-state analysis.

4. Published guidelines

Ayers and Westcot (1985) clearly described the procedure they used in developing their guideline. They assumed that the depth

Table 1

The average EC_e/EC_w as a function of the leaching fraction for four steady-state models for use in determining the leaching requirement. A&W refers to Ayers and Westcot (1985) and UC1 and UC2 refer to guidelines in Hanson et al. (2006).

| LF | A&W | UC1 | UC2 | Eq. (2) |
|------|-----|------|------|---------|
| 0.05 | 3.2 | 2.9 | 1.9 | 4.2 |
| 0.10 | 2.1 | 2.0 | 1.4 | 2.2 |
| 0.15 | 1.6 | | | 1.6 |
| 0.20 | 1.3 | 1.5 | 1.1 | 1.2 |
| 0.25 | 1.2 | | | 1.0 |
| 0.30 | 1.0 | 1.0 | 1.0 | 0.85 |
| 0.40 | 0.9 | 0.85 | 0.9 | 0.70 |
| 0.50 | 0.8 | 0.75 | 0.75 | 0.60 |

distribution of water uptake is 40, 30, 20, and 10% of total crop transpiration, corresponding with the first through the fourth quarter sections of the root zone, respectively. Using mass balance considerations, the salt concentration in the soil solution (EC_s) at the four quarter positions in the root zone were computed for LF values of 0.05 to 0.80. The linear average of these root zone salinities was divided by 2, thereby providing for an average root zone EC_e . The calculated average root zone EC_e for each value of LF can be used, as will be illustrated later, to determine the LR based on irrigation water salinity and crop salt tolerance.

Ayers and Westcot (1985) also presented the following equation developed by Rhoades (1974) as a guideline for calculating LR based on irrigation water salinity and crop salt tolerance.

$$LR = \frac{EC_w}{5EC_e^* - EC_w} \quad (2)$$

where EC_w is the electrical conductivity of the irrigation water.

Hanson et al. (2006) reproduced two graphs originally published by Rhoades (1999) showing the linear relationship between average root zone EC_e and EC_w for LF values between 0.05 and 0.50. One graph was for conventional surface and sprinkler irrigation (UC1) and the other for high frequency irrigation methods (UC2) such as drip.

A comparison of the four guidelines for calculating LR using these steady-state approaches is presented in Table 1. The numbers represent the ratio of the average root zone EC_e to EC_w , thus quantifying the salt concentrating factor associated with each listed value of LF. These ratio values can be used to assess maximum allowable irrigation water salinity values, such that the soil salinity concentration will not exceed the EC_e^* value for a specific crop (i.e. maximum crop yield). For example, consider a salt sensitive crop with an EC_e^* value of 1 dS/m that is irrigated with a LF of 0.15. The Ayers and Westcot (A&W) concentrating factor of 1.6 would indicate that the salinity of the applied water cannot exceed 0.62 dS/m for maximum yield. By comparison, the irrigation water salinity must be 0.31 dS/m or less to achieve maximum yield if the LF is 0.05. The A&W, UC1, and UC2 models predict that maximum yield will be achieved using a LF of 0.3 when the irrigation water salinity is equal to EC_e^* (Table 1).

The UC2 guideline has the lowest concentrating factors of the four guidelines, particularly at the lower values of LF. The UC2 guideline is based on high-frequency irrigation and a water-uptake weighted average EC_e (Rhoades, 1999), whereas the others are based on a linear average.

Ayers and Westcot (1985) reported in detail their analyses that led to the values of Table 1. Their analysis clearly demonstrates how root water uptake leads to a corresponding increase in soil salt concentration with depth, and to increasing soil salinity at a given depth as LF decreases. They show that the soil salinity values at the bottom of the rooting zone are greatly affected by LF values, and that the near-surface soil salinity is relatively least affected by the leaching magnitude. Their analysis clarifies that the method of

linearly averaging to determine the average root zone soil salinity is most sensitive to the salinity in the lower parts of the root zone, where less root water is extracted.

The assumption that plants respond to a linear-average soil root zone EC is not supported by experimental evidence. Gardner (1983) in his review of plant root water-extraction patterns concluded that soil water is initially mostly extracted from the near soil surface regions where plant roots are most prolific, and that the zone of root water-extraction progresses downwards as soil water becomes more limiting in the upper soil zones. Furthermore, there is experimental evidence that plants will extract additional soil water from the less-stressed portions of the root zone to compensate for reduced root water uptake in the stressed root zone regions. For example, van Schilfgaarde et al. (1974) analyzed data from an alfalfa experiment and concluded "apparently as long as the roots have access to water of low salinity, they are able to utilize some water of higher salinity without adverse effects." Based on what is scientifically known about water uptake by plants, it can be concluded a priori that determining yield from the linear average of salinity in the root zone will likely result in an overestimation of the negative impact of soil salinity on crop yield.

Given the analysis above, one may argue the presence of an inconsistency in our reasoning. On one hand, the M-H coefficients as reported are based on average root zone EC_e , yet the linear averaging of soil salinity by Ayers and Westcot (1985) is being considered as being at fault. This apparent contradiction is the result of the high leaching fractions used by the experiments analyzed by Maas and Hoffman (1977) to determine the salt tolerance levels. Under these conditions, the soil-water salinity is relatively constant with depth and is not concentrated greatly by plant water extraction. This condition differs largely from the high salt concentrations experienced in the deep root zones at low leaching fractions.

5. Shortcomings of the leaching requirement concept

Several shortcomings exist in the LR concept as has been proposed using steady-state analyses. The LR concept is based on achieving maximum yield. Maximum yields may not be the economically optimal yield. This is particularly true when only saline waters are available and potential maximum yields are not possible. Furthermore, there are situations, such as in the western San Joaquin Valley of California, where high costs are entailed in disposing subsurface drainage waters. Under these conditions there are economic benefits for reducing LF that must be considered in offsetting potential yield decreases. The relationships between water salinity, water application, crop yield, and the amount of drainage water are necessary to establish the optimal management strategy. The one-valued LR is not sufficient.

A steady-state analysis dictates that water is applied uniformly across the field at a constant rate and salinity. Clearly, such constant conditions are not realistic, but were required to obtain a simple leaching requirement solution. Rainfall or other forms of precipitation such as snow significantly moderates the consequences of irrigating with saline waters and evaluating its impact on crop yield is important. Only a few steady-state models consider rainfall. Moreover, crop rotations result in guidelines requiring differentiation between crops and years. Planting crops that vary in salinity tolerance is a common agricultural practice that cannot be incorporated into a steady-state analysis because of seasonal lag times associated with salinizing and desalinizing the soil profile. Leaching usually is not required for each irrigation event and this feature is not accommodated by steady-state analysis either.

Scheduling irrigation is most commonly based on accommodating potential ET (plant evapotranspiration) plus some extra for

nonuniform irrigation and leaching purposes. Few if any actually calculate the amount of extra water needed to achieve a desired LF. Some assume that a 20% LF is achieved by applying 20% more water than ET. That is erroneous. The relationship between applied water (AW), ET and LF is

$$\frac{AW}{ET} = \frac{1}{1 - LF} \quad (3)$$

A LF equal to 0.20 requires AW/ET to be 1.25. Thus 25% more water than ET must be applied to get a 20% LF. A LF of 50% is achieved by applying twice the amount of ET. Potential ET (PET) is defined as the maximum daily or seasonal total plant ET, implying zero crop stress as caused by either reduced irrigation water application or by soil salinity. In Eq. (3), ET represents actual crop evapotranspiration. However, if defined in the context of computing irrigation LR with maximum crop yield, in fact, ET in Eq. (3) would be PET.

In addition, micro-irrigation systems typically apply water locally, thus wetting only part of the field soil surface. Consequently, field-wide application of Eq. (3) leads to underestimation of actual LF values. For example, Hanson et al. (2009) demonstrated that high efficiency of drip micro-irrigation that partially wets the soil surface area occurred only under conditions of severe deficit irrigation as determined from a field-scale water balance of tomatoes under shallow groundwater table conditions. Hence, LR guidelines must be adjusted, to allow for partial soil surface wetting and spatially variable salt accumulation.

6. Transient-state analyses

The advent of high-speed computers has facilitated the opportunity to develop models based on transient-state analyses. These models allow simulations that include temporal changes in crop, changes in crop salt tolerance through the growing season, water salinity including rain, and the amount of irrigation and rain that are consistent with actual conditions. Several models have been published in the literature. The workgroup developed a matrix to compare various features of the models. Some models have several features that are the same, but each has at least one component that differs from the others.

Models are only useful to the extent that they accurately simulate behavior in the field. Comparison between model simulations and experimental data from field experiments is important. Unfortunately because of the complexity and cost associated with field experiments that include both water salinity and other variables, not many field experiments have been conducted. Among the few, one was conducted at the Gilat Agricultural Experimental Station in the northern Negev of Israel on corn (Shalhevet et al., 1986). Five irrigation water salinities ranging from 1.7 to 10.2 dS/m were used along with four irrigation intervals ranging from 3.5 to 21 days. Feng et al. (2003) compared the experimentally measured yield to the simulated yields using the ENVIRO-GRO model. The mean simulated relative yield was 0.70 and the measured mean relative yield was 0.68. The Willmott's index of agreement between simulated and measured yield was 0.96 (a value of 1.0 represents perfect agreement). Based on these results, the authors concluded that the ENVIRO-GRO model can be used with confidence in simulating the consequences of irrigation management options under saline conditions.

Another frequently used transient model is HYDRUS-2D, simulating movement of water and dissolved salts in soils. This model was used by Hanson et al. (2008) to compare results with field experiments on processing tomatoes under shallow water table conditions for both drip and sprinkler irrigation for a wide range of irrigation water salinities. Replicated experiments were conducted to investigate relationships among yield, irrigation water salinity and applied water. Both field and model results showed

Table 2

The AW/PET value calculated with ENVIRO-GRO (E-G) and several steady-state models to achieve maximum corn yield when irrigating with water of 1 and 2 dS/m. A&W refers to Ayers and Westcot (1985), UC1 and UC2 refer to guidelines in Hanson et al. (2006), and E-G refers to ENVIRO-GRO (Pang and Letey, 1998).

| EC _w | A&W | UC1 | UC2 | Eq. (2) | E-G |
|-----------------|------|------|------|---------|-------|
| 1 | 1.19 | 1.19 | 1.11 | 1.15 | <1.05 |
| 2 | 1.82 | 1.67 | 1.75 | 1.44 | 1.17 |

that seasonal field-wide water applications should be about equal to seasonal ET, thereby providing adequate localized leaching at the crop root zone scale and preventing the shallow saline water table from rising.

7. Comparison of results between transient-state and steady-state analyses

Letey and Feng (2007) compared results of transient-state analysis using ENVIRO-GRO (Pang and Letey, 1998) with steady-state analyses for irrigating corn. The results presented in that paper are expanded here to include all of the steady-state models that are reported in Table 1. Transient-state models do not directly calculate LR. Instead, model simulations are conducted for a series of seasonal water applications, from which the lowest application is selected that maintains maximum crop yield, so that ET in Eq. (3) is PET. Thus the input is the ratio of AW to PET. Eq. (3) was used to convert LR to AW/PET or vice-versa. The results are presented in Table 2 for irrigating corn with water salinity values of 1 or 2 dS/m. The EC_e* for corn was assumed to equal 1.7 dS/m (Maas and Hoffman, 1977). All steady-state methods predicted that more water had to be applied to achieve maximum yield as compared to ENVIRO-GRO. The differences were especially great at the higher water salinity.

Corwin et al. (2007) also reported that the calculated LR was lower when determined using a transient-state approach than when using a steady-state approach. They used TETrans and UNSATCHEM as their transient-state models. They calculated that the reduced LR using the transient-state analyses as compared to the commonly used traditional method for the Imperial Valley of California would result in a diminished drainage volume of approximately 125 Mm³.

Some of the reasons that the transient-state simulations resulted in a lower required irrigation amount than the steady-state approaches to achieve maximum yield are as follows:

- (1) The steady-state approach assumes that the plant responds to the average root zone salinity that increases greatly as the LF decreases. However, the major amount of water is extracted from the upper parts of the root zone where the salt concentration is not very sensitive to LF.
- (2) The salt concentration at a given depth in the field does not remain constant with time as predicted by the steady-state models, but is continually changing as water is added or extracted by the plant. The salts become concentrated by water extraction, but irrigation water “flushes” the salts downwards and thus reduces the concentration to a lower value at a given depth after irrigation. The concentration immediately after irrigation near the soil surface would be close to the concentration in the irrigation water.
- (3) For most soils, the volumetric soil–water content would be reduced by less than half between irrigations; thus, concentrating the soil water salinity by less than two between irrigations. Therefore, as a general guideline, irrigation water with a salt concentration equal to EC_e* can be used with a relatively low LF. This conclusion is based on the fact that the M–H coefficients are based on EC_e which is about EC_s/2. The soil–water can there-

fore be concentrated by a factor of two without exceeding the threshold value.

Chemical factors also impact the salinity in the root zone. Salt precipitation will affect the soil-solution salinity and the salt precipitating will be a function of irrigation water chemistry and soil chemistry. Precipitation of calcite occurs in almost all instances when irrigating in arid regions at low leaching fractions. This typically reduces the salt load by 5–25% at the bottom of the root zone and reduces the EC_s throughout all but the upper portion of the root zone. For waters containing elevated sulfate and calcium concentrations, precipitation of gypsum will cause a major reduction in root zone EC_s . An additional factor reducing the EC_s to below that predicted based on water budgets (as above) is the non-linearity between EC of water and the actual concentration (the EC/salt concentration ratio decreases with increasing concentration). Additionally, if field measurements of soil EC_e are made, and related to the salt tolerance tables, salinity impacts will again be overestimated. In this instance mineral dissolution (primarily gypsum if present) will release salts in the relatively dilute saturation extract.

All of the analyses presented here ignored rainfall. Rainfall at any time of the year would partially mitigate the impacts of irrigating with saline waters. Simulations using transient-state models for actual field conditions require input of all waters, including rain. Thus the impact of rain cannot be ignored using these models. The effect of rain can be estimated assuming quasi steady-state conditions and using the weighted-average salinity of combined rain and irrigation water. However, the other deficiencies of the steady-state analyses are not corrected by this procedure. For example in a model developed by Isidoro and Grattan (2010), simulations indicate that rainfall occurring mostly in the winter produced lower EC_e values in the root zone than did rainfall evenly distributed throughout the year. This result suggests that temporal distributions of rain may have an important influence of seasonal root zone salinity.

8. Conclusions

The workgroup concludes that the present guidelines based on steady-state analyses overestimate the leaching requirements and the negative consequences of irrigating with saline waters. This error is particularly large at low leaching fractions. This is a fortuitous finding because irrigating to achieve low leaching fractions

is desirable for the purpose of reducing the transport of chemicals that degrade groundwater quality and also provides for a more efficient use of limited water supplies. The feasibility of using saline waters for irrigation is also enhanced. Thus these positive goals can be pursued without an erroneous overestimate of developing soil salination. However, soil salination is still a potentially very negative consequence of irrigation and cannot be ignored.

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